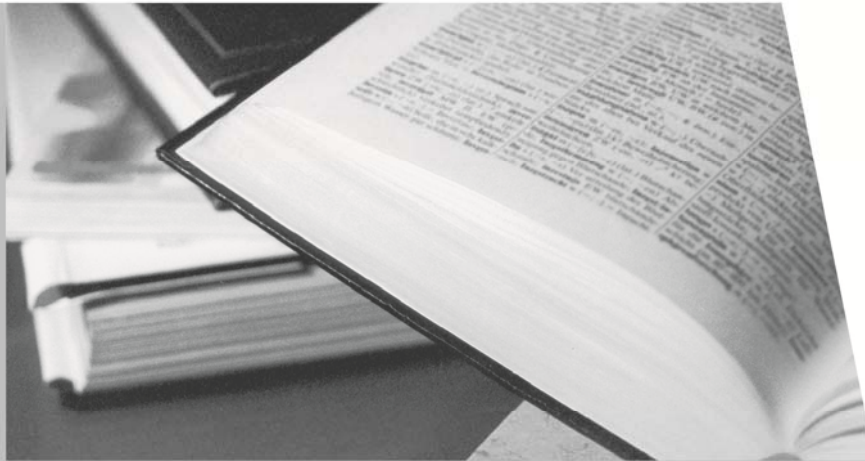


# Working Papers



The role of urban green space for  
human well-being  
by Christine Bertram and Katrin Rehdanz

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## **The role of urban green space for human well-being\***

Christine Bertram and Katrin Rehdanz

**Abstract:**

Most people in Europe live in urban environments. For these people, urban green space is an important element of well-being, but it is often in short supply. We use self-reported information on life satisfaction and different individual green space measures to explore how urban green space affects the well-being of the residents of Berlin, the capital city of Germany. We combine spatially explicit survey data with spatially highly disaggregated GIS data on urban green spaces. We observe a significant, inverted U-shaped effect of the amount of and distance to urban green space on life satisfaction. According to our results, the optimal amount of green space in a 1 km buffer is 36 ha, or 11.5% of the buffer area, and 75% of the respondents have less green space available. Our results are robust to a number of robustness checks.

**Keywords:** life satisfaction, urban ecosystem services, urban green space, well-being

**JEL classification:** I31, Q51, Q57, R00

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## 1 Introduction

Approximately 75% of Europeans live in urban areas (World Bank, 2013). One important element for their well-being and quality of life is the availability of urban green space. There are different ways in which urban green space can positively influence well-being and health (see Tzoulas et al. (2007) for an overview). Benefits can accrue from increased activity levels as a result of being in contact with nature (see Bowler et al. (2010) for a review). In addition, Kaplan (2001) shows that natural elements in the view from a window can have positive effects. Further benefits are brought about by the moderation of adverse environmental conditions such as air pollution, high temperatures, or noise (e.g. Gidlöf-Gunnarsson and Öhrström, 2007). However, in most urban areas, and particularly in inner-city areas, green spaces are in insufficient supply (Kabisch and Haase, 2011).

Individual countries and/or cities have begun to take an increasing responsibility in developing urban green space and improving the services provided by different forms of urban green. Following the Convention on Biological Diversity (CBD; UN, 1992), these countries and cities have formulated national, regional, or local action plans to integrate urban biodiversity and ecosystem services (ES) provided by urban green space, among others, into management.<sup>1</sup> The German National Strategy on Biological Diversity, for example, calls for an increase in green space in settlement areas (BMU, 2007).<sup>2</sup> At the city level, some German cities have defined minimum targets for per capita supply of urban green space.<sup>3</sup>

City development, however, always has to address trade-offs and conflicting interests between inner city development, e.g., for housing, and the development or preservation of green and open spaces (Schetke et al., 2012). Information on the benefits and costs of alternative land uses can therefore be valuable in supporting decision-making and ensuring that land is used sustainably, meeting the needs of the inhabitants.<sup>4</sup> Information on the benefits of ES, which are mostly not traded on markets, however, is often not available. Despite the relevance of urban ES for a large share of the population, environmental valuation studies have so far focused on ES in rural contexts. Existing studies on the economic valuation of urban green space have mostly used traditional techniques such as stated or revealed preference approaches (see Brander and Koetse (2011) and Perino et al. (2013) for recent meta-analyses).

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<sup>1</sup> See TEEB (2011) for an overview of the ES concept and its application in an urban context.

<sup>2</sup> This objective is integrated into federal law by requiring that open spaces in urban and peri-urban areas have to be preserved and developed where they are not sufficiently available (§ 1 Abs. 6 BNatSchG, 2009).

<sup>3</sup> The City of Berlin, e.g., has the goal to provide 6m<sup>2</sup> of public green space per inhabitant (SSUB, 2013).

<sup>4</sup> The EU Biodiversity Strategy (EC, 2011), for example, requires all member states to assess their ecosystems and the economic value of those systems by 2020.

A recent alternative in the field of environmental valuation is the life satisfaction approach (LSA).<sup>5</sup> The two existing economic studies analyzing the effect of urban green on life satisfaction cover cities in Australia (Ambrey and Fleming, 2013) and China (Smyth et al., 2008). Unlike Ambrey and Fleming (2013) and Smyth et al. (2008), we use an individual green space measure that captures the area of green space surrounding a respondent's home. We further add to this literature by exploring and comparing different ways in which urban green might affect the well-being of city inhabitants. Moreover, we offer the first application of the LSA to value urban green in a European city, namely Berlin, the capital city of Germany.

Berlin, located in the Eastern part of Germany with an area of 892 km<sup>2</sup> (SSUB, 2012a) and a population of 3.4 million (ASBB, 2013a), is particularly interesting. The expected population growth and the trend toward smaller household sizes will exert strong pressure on existing green spaces in the inner-city districts, particularly if a densification strategy is to be followed and urban sprawl is to be avoided. Such a conflict can currently be observed in the case of the Tempelhofer Feld.<sup>6</sup> On the other hand, there are still many open spaces such as brownfields that might be turned into residential or commercial areas (Simons et al., 2012).

The objective of this article is to answer the following research questions: (1) In which ways, if any, does urban green space affect the well-being of people? (2) Is more green space always better, or is there an optimal level of urban green? (3) What is the monetary equivalent of a change in the availability of urban green space?

To address these questions, we use spatially explicit survey data of Berlin residents together with spatially highly disaggregated GIS data on urban green spaces. We use several indicators for our analyses. Based on land cover data from the Urban Atlas (EEA, 2012), we calculate the amount of green space available in the living environment of each respondent as well as the distance to the nearest green space. Based on our survey data, we analyze the frequency of park visits and a dummy variable indicating whether the respondent has a view of a park from his/her home.

The paper is structured as follows. Section 2 provides a review of the literature on the economic valuation of urban green and the literature on economics and subjective well-being. Section 3 presents the empirical approach and the data. Section 4 reports the results of the main regressions and sensitivity analyses. Section 5 discusses the results and presents the conclusions.

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<sup>5</sup> Self-reported life satisfaction is used as a proxy for subjective well-being. Please note that we use the terms life satisfaction and well-being interchangeably throughout the paper.

<sup>6</sup> The so-called Tempelhofer Feld is located on the area of the former airport Berlin Tempelhof. The associated free areas including the former airfield have a size of 303 ha. Currently, it can be used by the public, e.g., for recreational purposes and is left more or less in its original state (GrünBerlin GmbH, 2013). There are conflicting interests concerning the future development of the area.

## **2 Economic valuation of urban green space – literature review**

### **2.1 Stated and revealed preference methods**

Despite the relevance of urban ES for city inhabitants, there are few economic studies that elicit the value of urban ES using either stated preference methods such as contingent valuation (CV) or choice experiments (CE) or revealed preference methods such as hedonic pricing (HP) or travel costs (TC). Even fewer environmental valuation studies specifically focus on urban green spaces or parks.

The results of a range of CV and HP studies are analyzed in two recent meta-analyses that focus on different types of urban ecosystems and have different regional foci. Brander and Koetse (2011) provide a meta-analysis of 32 international CV and HP studies valuing urban open spaces with a focus on the USA. They find that most of the CV studies refer to urban forests and urban agriculture and much fewer studies investigate urban green spaces and parks. HP studies, in contrast, mostly investigate the role of urban parks and green spaces for property prices. A more recent meta-analysis for the UK is provided by Perino et al. (2013). It is based on five studies analyzing the effect of increased distance to formal recreation sites and city-edge green space on property prices using HP, CV, and expert interviews.

With respect to CE, there are even fewer examples of studies analyzing preferences for urban ES. The only study that values urban green spaces or parks that we are aware of is an application for Dublin, Ireland, by Bullock (2006).<sup>7</sup> Two examples of TC studies are Fleischer and Tsur (2003), who use the individual TC method to estimate the economic value of urban parks in Israeli cities, and Chaudhry and Tewary (2006), who use zonal TC to assess the recreational value of urban forests in Chandigarh, India.

### **2.2 The life satisfaction approach**

The LSA is a recent alternative in the field of environmental valuation. It is based on the assumption that environmental (dis)amenities are among the factors that determine subjective well-being (SWB). Following this approach, self-reported life satisfaction is taken as a proxy for SWB and estimated as a function of factors such as environmental amenities and income, while at the same time controlling for other socio-economic, demographic, and geographical information. Based on the assumption that life satisfaction data are an approximation of what Kahnemann et al. (1997) labeled “experienced utility,” this estimated relationship is used to derive the implicit marginal rate of substitution (MRS) between income and the environmental amenity in question.

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<sup>7</sup> Examples of CE used in other urban contexts are studies by Lanz and Provins (2013), who focus on local environmental improvements in the UK, or Bae (2011), who analyzes preferences for urban stream restoration in Korea.

Unlike stated preference methods, this method does not ask people to place a monetary value on a complex environmental good in a hypothetical situation. Compared to CV, this may reduce biases resulting from the hypothetical nature of the decision and from potentially strategic behavior. In comparison to the HP method, the LSA does not rely on decisions being reflected in actual market transactions. Thus, it is not affected by biases resulting from the assumption that the housing market is in equilibrium, which it often is not. Moreover, biases from distorted risk perceptions may play a lesser role. There are, however, also limitations to the LSA. One precondition is that life satisfaction data, which are used as a proxy for well-being or utility, satisfy appropriate quality requirements (being ordinal in character, consistent, valid, and reliable).<sup>8</sup> For a discussion of the underlying assumptions and implications of the LSA in comparison with CV and HP, see Frey et al. (2010).

Research on SWB has identified a number of personal, demographic and socio-economic factors that explain differences in SWB (see, e.g., Dolan et al. (2008) for an overview). With regard to environmental (dis)amenities, most of the studies have looked at air pollution. The most recent examples are studies by Levinson (2012), Ferreira and Moro (2010), Luechinger (2009, 2010), MacKerron and Mourato (2009), and Rehdanz and Maddison (2008). Other environmental issues investigated include climate (Ferreira and Moro, 2010; Brereton et al., 2008; Rehdanz and Maddison, 2005; Frijters and van Praag, 1998), noise (Rehdanz and Maddison, 2008; van Praag and Baarsma, 2005), scenic amenities (Ambrey and Fleming, 2011), protected areas (Ambrey and Fleming, 2012), land cover (Kopmann and Rehdanz, 2013), droughts (Carroll et al., 2009), and floods (Luechinger and Raschky, 2009).

Many of the earlier environmental studies look at nationwide or cross-country data sets and suffer from a lack of more disaggregated environmental data (e.g., Welsch, 2006 or Rehdanz and Maddison, 2005) and are thus not able to take more disaggregated spatial controls into account. Some studies do include spatial controls, e.g., accounting for the fact of whether people live in urban, rural or peri-urban areas (e.g., Ferreira and Moro, 2010). Few studies explicitly address urban environments or data sets customized to urban environments. One exception is MacKerron and Mourato (2009), who look at air quality in London using spatially disaggregated data.

Two studies investigate the amenity value of urban green spaces. Using wave 5 of the HILDA survey, Ambrey and Fleming (2013) investigate the role of public green space for the well-being of people in major Australian cities. The green space measure they use is the percentage of public

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<sup>8</sup> See Welsch and Kühling (2009) for a discussion of conceptual and methodological issues. In Section 5, we address additional issues potentially relevant to this analysis.

green space in the resident's collection district.<sup>9</sup> The estimated implicit MRS for a 1% (equivalent to 143 m<sup>2</sup>) increase in public green space is AUD 1,168 in terms of annual household income. Smyth et al. (2008) use survey data gathered from the inhabitants of 30 Chinese cities to estimate the effects of pollution, disasters, congestion and green space on human well-being. The green space measure they use is the area of green space per capita on city-level. They find a statistically significantly positive effect of green space on life satisfaction for the model specification with city dummy variables. However, MRS estimates are not reported and cannot be derived.

### 3 Methodology and data

#### 3.1 Methodological approach and empirical strategy

We estimate the effects of different demographic, socio-economic, and environmental variables on life satisfaction using the following regression equation:

$$LS_{ij} = \beta_0 + \beta_y \ln(Y_i) + \beta_x X_i + \beta_z Z_i + \beta_s S_j + \beta_a A_i + \varepsilon_{ij} \quad (1)$$

The dependent variable,  $LS_{ij}$ , is the stated life satisfaction of respondent  $i$  living in district  $j$ . Explanatory variables include  $Y_i$ , which is the individual net monthly income of respondent  $i$ .<sup>10</sup> Income enters the regression equation in its natural logarithm to account for the declining marginal utility of income. Further explanatory variables are captured in the vectors  $X_i$  and  $Z_i$ , which contain other demographic and socio-economic characteristics of respondent  $i$  and of her household, respectively (see Table A-1 in Appendix A for a description and summary statistics of all variables). The variable  $S_j$  contains dummy variables for each district  $j$  to control for district-specific effects, and  $\varepsilon_{ij}$  is the error term.

We estimate a set of different specifications that differ by the way the environmental variable  $A_i$  is measured. We analyze i.) the amount of urban green space available in the living environment of respondent  $i$ , ii.) the distance to the nearest urban green space bigger than 5 ha, iii.) the frequency of visits to urban parks, and iv.) whether respondent  $i$  has a view of a park from his/her home. The indicators for a park view and the number of park visits are based on self-reports derived from our

<sup>9</sup> The collection district is the smallest spatial unit in the Australian Standard Geographical Classification. Assuming each collection district takes the shape of a circle, the median radius from the centroid is approximately 750m (Ambrey and Fleming, 2013).

<sup>10</sup> Total net monthly individual income was calculated by dividing the corresponding household income by the weighted number of household members according to the OECD-modified equivalence scale (OECD, 2009). Because there was no information about children under the age of 14 in the household in the survey, this was adapted to children under the age of 12. Household income was indicated in ranges. We used the midpoint of the indicated range to calculate the corresponding individual income. The use of midpoints is common in life satisfaction studies if income is given in ranges (Carroll et al., 2009; MacKerron and Mourato, 2009).

survey. The amount of urban green space and the distance to the nearest urban green space are calculated based on the residential address of the respondents and land cover data from the Urban Atlas (EEA, 2012).<sup>11</sup>

In addition to the linear regression model described in equation (1), we also estimate a non-linear form in which the amount of and distance to green spaces enter the regression in their linear and in their squared form. In the non-linear case, the estimation equation changes to

$$LS_{ij} = \beta_0 + \beta_y \ln(Y_i) + \beta_x X_i + \beta_z Z_i + \beta_s S_j + \beta_{a1} A_i + \beta_{a2} A_i^2 + \varepsilon_{ij} \quad (2)$$

This functional form presumes that the effect of urban green may also depend on its current allocation. The parameter  $\beta_{a1}$  is expected to be positive, while  $\beta_{a2}$  is expected to be negative.

The estimated relationships can then be used to derive the implicit MRS between the environmental variable, e.g., the individually available amount of urban green space,  $A_i$ , and individual income,  $Y_i$ , by dividing the absolute value of the derivative of life satisfaction ( $LS_{ij}$ ) with respect to  $A_i$  by the derivative of  $LS_{ij}$  with respect to  $Y_i$ . For the linear specification, the MRS, evaluated at the mean of income, can thus be calculated as follows:

$$MRS = \frac{\partial LS_{ij} / \partial A_i}{\partial LS_{ij} / \partial Y_i} \Bigg|_{dLS_{ij}=0} = \frac{\beta_a}{\beta_y} \bar{Y} \quad (3)$$

For the non-linear specification, the current allocation of the environmental good has to be taken into consideration. The implicit MRS between  $A_i$  and  $Y_i$  evaluated at the means of the environmental variable and income can be calculated as follows:

$$MRS = \frac{\partial LS_{ij} / \partial A_i}{\partial LS_{ij} / \partial Y_i} \Bigg|_{dLS_{ij}=0} = \frac{\beta_{a1} + 2 \beta_{a2} \bar{A}}{\beta_y} \bar{Y} \quad (4)$$

<sup>11</sup> See section 3.2 for a more detailed description of these data.



## 3.2 Data

### 3.2.1 Survey data

Most of the data used for our analysis come from a web survey carried out in Berlin, Germany, in September 2012.<sup>12</sup> The main objective of this survey was to investigate the role of urban green space, particularly parks, for the well-being of people living in urban environments. For this purpose, the survey included a number of questions on park use patterns and the perception of the environment. The survey also included questions on socio-economic and demographic characteristics of the respondents and their households, including gender, age, marital status, education, occupation, and household income. To take account of the fact that the housing environment is very important for personal well-being in urban surroundings, we also included questions on the housing type and on neighborhood characteristics in the survey (see Table A-1 in Appendix A for summary statistics).

Survey participants were screened to ensure that they had been living in Berlin for more than one year. Only residents of the districts of Mitte, Kreuzberg-Friedrichshain, Pankow, Charlottenburg-Wilmersdorf, Tempelhof-Schöneberg, Neukölln and Lichtenberg were included in the survey (see Figure 1). These districts were selected because they include densely populated inner-city districts of Berlin with a relatively homogenous distribution of green space and exclude districts with large shares of water areas and forests. The districts were also selected to be comparable with the whole population of Berlin with regard to age and gender. Moreover, a balanced distribution between formerly Eastern and Western German parts of the city of Berlin has been targeted. The final sample consists of 485 usable observations.

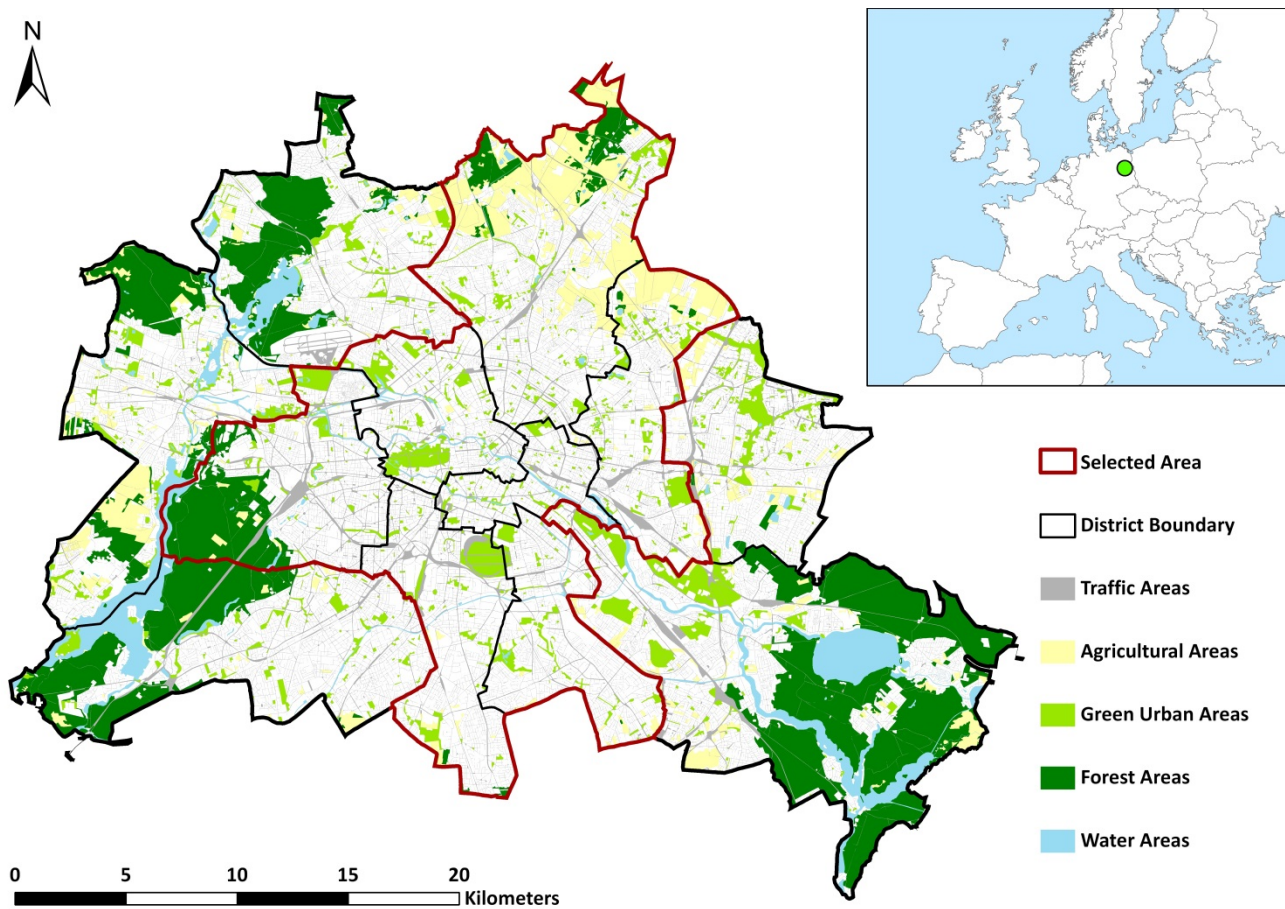
The question about subjective well-being was phrased as follows: “All things considered, how satisfied are you with your life these days?” Respondents were asked to answer the question on an 11-point Likert scale ranging from “0” (very dissatisfied) to “10” (very satisfied). This question mode is in line with many large surveys carried out in different countries and used in the economic literature on life satisfaction (Welsch and Kühling, 2009). The mean value for life satisfaction in our sample was 6.8 with a standard deviation of 2.05, which is comparable to other life satisfaction studies for Germany (e.g., Kopmann and Rehdanz, 2013).

The distribution of gender in our sample is comparable to that in Berlin and in the selected districts of Berlin. Regarding age, people between 50 and 64 are overrepresented by approximately 15% in our sample, while people above 64 are underrepresented to the same extent. This, however, seems

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<sup>12</sup> More detailed information about the survey is provided in Appendix B.

unavoidable, given that we use a web survey. The distribution of income in our sample is comparable to that in Berlin and in the selected districts.



**Figure 1. Distribution of urban green spaces and other natural and semi-natural areas in the city of Berlin. Own presentation based on Urban Atlas data (EEA, 2012).**

As indicated in section 3.1, two of the environmental indicators we use to analyze the well-being effects of urban green are derived from the survey, namely the indicators of the “frequency of park visits” and “park view.” The “frequency of park visits” is an ordinal variable measured on a 6-point scale, ranging from “never” to “(almost) daily.”<sup>13</sup> The dummy variable “park view” is derived from a question asking respondents whether they have a view of a park from their homes. Approximately 14% of the respondents report having such a view.

We use self-reported information on residential addresses from the survey to locate respondents in the city of Berlin in a spatially explicit way using ArcGIS. This enables us to link the survey data with GIS data on urban green spaces. Out of the 485 usable observations, 11% provided their full address, including street name and number. Table 1 gives an overview of the type of information we

<sup>13</sup> Frequencies were averaged over the stated number of visits in the summer and winter of the 12 months preceding the survey.

gathered from the respondents, the way we located them in ArcGIS, and the location statuses we assigned to them based on a decreasing level of precision. Given that we are interested in questions relating very specifically to the living environment of the respondents, we decided to limit the subsequent analyses to those respondents that indicated at least the name of the street (location statuses 1 and 2).

**Table 1. Overview of different location statuses.**

Location status	Description	Location in ArcGIS	Number of observations	Percentage
1	Full address	Precise address	53	7.2%
2	Street without number	Middle of the street	326	67.2%
3	ZIP code	Centroid of the ZIP code area <sup>a</sup>	50	10.3%
4	District	Centroid of the district <sup>b</sup>	56	11.5%

<sup>a</sup> The average area of the ZIP code areas in which the respondents live is 3 km<sup>2</sup>.

<sup>b</sup> The average area of the selected districts is 54 km<sup>2</sup>.

### 3.2.2 GIS data and social indicators

The spatial data on urban green spaces are taken from the Urban Atlas (EEA, 2012). This database provides pan-European comparable land use and land cover data for large urban areas with more than 100,000 inhabitants on a 50x50 m grid level (EU, 2011). We use the areas designated as “green urban areas” for our analyses.<sup>14</sup>

To calculate the individual availability of urban green space, we created buffers with a diameter of 1 km around the respondents’ residential addresses and calculated the amount of green space within the buffer area for each respondent. This gives us an individual measure of green space availability on the respondent level. However, we have to be aware of potential measurement errors in the green space variable due to the uncertainties related to the location of the respondents described in section 3.2.1. Thus, we excluded observations in the 1st and 99th percentile of the variable “green space” from our sample to mitigate the influence of outliers. This led to an exclusion of 10 observations from the sample. The total buffer area for the buffer with a 1 km radius is 314.2 ha (or

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<sup>14</sup> The land use category “green urban areas” includes public green areas for predominantly recreational use such as gardens, zoos, parks, or castle parks. Not included are private gardens within housing areas, cemeteries, buildings within parks, such as castles or museums, patches of natural vegetation or agricultural areas enclosed by built-up areas without being managed as green urban areas. We also include the lawns belonging to the former Tempelhof airport, which are now used for recreation, in the measure for urban green space. This was not the case in the original data set (EU, 2011).

approximately 3.14 km<sup>2</sup>). Individual green space availability is between 0.5% and 31.2% of this buffer area, i.e., between 1.1 ha and 98.1 ha.<sup>15</sup>

Based on the same Urban Atlas data set, we also calculated the distance from each respondent's individual residential address to the nearest urban green space greater than 5 ha. The mean distance is 661 m, ranging from a minimum of 0.5 m to a maximum of 2,495 m.

In addition to land cover data from the Urban Atlas, we use sealing data provided by the City of Berlin for a sensitivity analysis (SSUB, 2012c; see also section 4.3.4). Sealing data do not only reflect the supply of public green spaces but also the existence of other types of public green or private green such as gardens. High degrees of sealing thus often go hand in hand with a mismatch between the number of inhabitants and the supply of green spaces and imply reduced regulating services such as local climate regulation. The GIS data set provides average degrees of sealing in Berlin on the block level for the year 2011.<sup>16</sup> We create the variable "sealing" by calculating the area-weighted average degree of sealing in a buffer area around the respondents' residential addresses.

In addition to environmental data, we use highly disaggregated data on the social status of neighborhoods for additional robustness checks (see section 4.3.1). The City of Berlin publishes social indicators such as different measures of unemployment on the level of planning units, which are considerably smaller than the districts and may thus give a good picture of the local social status of a neighborhood (SSUB, 2011).<sup>17</sup>

#### **4 Estimation results**

Results of the regressions estimating the effect of several explanatory variables on life satisfaction using ordered logit models are shown in Table 2.<sup>18</sup> We cluster error terms on the district level to calculate robust estimates as we combine variables on the individual level and on the district level in our regressions (Moulton, 1990). Model 1 presents the results of a baseline specification focusing on demographic and socio-economic variables. Models 2 to 7 additionally include environmental variables to estimate their effect on life satisfaction. We add the environmental indicators to the regressions one by one because there is a significant correlation between them.

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<sup>15</sup> For comparison, the area of the selected districts in Berlin is between 2,034ha in the case of Friedrichshain-Kreuzberg and 10,315ha in the case of Pankow. The area share of public green space is between 6.8% in Pankow and 14.8% in Mitte (SSUB, 2012a).

<sup>16</sup> Blocks have an average size of 3.6 ha.

<sup>17</sup> Planning units have an average size of 2 km<sup>2</sup> and are thus approximately two-thirds the size of the 1 km buffers used to calculate the availability of urban green spaces.

<sup>18</sup> Ordered logit is used because life satisfaction is measured on an ordinal scale. In accordance with the literature (e.g., Ferreira and Moro, 2010 or Brereton et al., 2008), we find that OLS yields very similar results compared to ordered logit.

**Table 2. Results of main regressions (ordered logit).**

Life satisfaction ( $LS_{it}$ )	Model 1	Model 2	Model 3	Model 4	Model 5	Model 6	Model 7
<u>Individual demographic and socio-economic characteristics (<math>Y_i, X_i</math>)</u>							
Log individual income	0.8577*** (0.28)	0.8545*** (0.27)	0.8332*** (0.27)	0.8686*** (0.28)	0.9020*** (0.30)	0.8247*** (0.26)	0.9720*** (0.27)
Gender	-0.0465 (0.19)	-0.0409 (0.20)	-0.0852 (0.22)	-0.0719 (0.20)	-0.0431 (0.21)	-0.0427 (0.23)	-0.0578 (0.22)
Age	-0.1091** <sup>b</sup> (0.04)	-0.1081*** <sup>b</sup> (0.04)	-0.1093** <sup>b</sup> (0.04)	-0.1172** <sup>c</sup> (0.05)	-0.1215*** <sup>b</sup> (0.04)	-0.1151*** <sup>b</sup> (0.04)	-0.1112** <sup>c</sup> (0.05)
Age squared	0.0013*** <sup>b</sup> (0.00)	0.0013*** <sup>b</sup> (0.00)	0.0013*** <sup>b</sup> (0.00)	0.0014** <sup>c</sup> (0.00)	0.0015*** <sup>b</sup> (0.00)	0.0014*** <sup>b</sup> (0.00)	0.0013** <sup>c</sup> (0.00)
Single	Reference	Reference	Reference	Reference	Reference	Reference	Reference
Married	0.5032*** (0.12)	0.5009*** (0.11)	0.5273*** (0.14)	0.5119*** (0.11)	0.5827*** (0.14)	0.5353*** (0.12)	0.4809*** (0.15)
Partner	0.4137* (0.24)	0.4082 (0.25)	0.4002** (0.20)	0.4216* (0.23)	0.4638** (0.20)	0.4445** (0.21)	0.3955* (0.23)
Separated, divorced or widowed	0.0290 (0.44)	0.0246 (0.44)	0.0902 (0.44)	0.0516 (0.41)	0.1048 (0.43)	0.0663 (0.37)	-0.0160 (0.44)
Bad or very bad health	Reference	Reference	Reference	Reference	Reference	Reference	Reference
Fair health	1.0441*** (0.26)	1.0445*** (0.26)	1.1657*** (0.24)	1.0619*** (0.24)	1.0716*** (0.23)	1.0781*** (0.25)	1.1328*** (0.27)
Good health	2.2097*** (0.36)	2.2083*** (0.36)	2.3672*** (0.36)	2.2728*** (0.36)	2.3798*** (0.42)	2.2299*** (0.35)	2.2868*** (0.33)
Very good health	3.4648*** (0.27)	3.4599*** (0.26)	3.5843*** (0.27)	3.4843*** (0.25)	3.5538*** (0.21)	3.3166*** (0.28)	3.5301*** (0.24)
Basic education	Reference	Reference	Reference	Reference	Reference	Reference	Reference
Secondary education	1.3072*** (0.49)	1.3182*** (0.51)	1.3730*** (0.52)	1.2554*** (0.45)	1.3027*** (0.42)	1.2663** (0.53)	1.2805*** (0.46)
Tertiary education	1.2405** (0.49)	1.2548** (0.49)	1.3387*** (0.52)	1.1587** (0.46)	1.1176** (0.47)	1.1916** (0.49)	1.2267*** (0.38)
Full-time employed	Reference	Reference	Reference	Reference	Reference	Reference	Reference
Part-time employed	0.8963** (0.37)	0.9092** (0.39)	0.8510** (0.39)	0.8360** (0.35)	0.8637** (0.35)	0.8335** (0.34)	0.8857** (0.36)
Unemployed	-0.3141 (0.87)	-0.3158 (0.87)	-0.3410 (0.87)	-0.3048 (0.84)	-0.3617 (0.79)	-0.3178 (0.92)	-0.1367 (0.73)
Unable to work	1.0080 (1.05)	1.0066 (1.05)	1.0547 (0.98)	0.9662 (1.00)	0.9603 (1.05)	1.1763 (1.11)	1.1175 (0.95)
Retired	0.1599 (0.79)	0.1684 (0.81)	0.1264 (0.79)	0.1374 (0.79)	0.1257 (0.78)	0.1133 (0.80)	0.1591 (0.81)
Student	-0.0033 (0.43)	0.0074 (0.42)	0.0438 (0.46)	-0.0590 (0.37)	0.0455 (0.40)	-0.0612 (0.42)	0.0512 (0.38)
Other occupation	-0.5159 (0.52)	-0.4974 (0.51)	-0.5365 (0.48)	-0.6407 (0.55)	-0.6504 (0.52)	-0.5500 (0.50)	-0.4538 (0.48)
Migration background	-0.0483 (0.24)	-0.0483 (0.24)	-0.0022 (0.28)	-0.0386 (0.25)	-0.0662 (0.25)	-0.0570 (0.25)	-0.0088 (0.26)
Lifetime	0.9514* (0.49)	0.9571* (0.49)	0.9459** (0.43)	0.8911* (0.47)	1.0202** (0.46)	0.9958** (0.46)	0.7730* (0.42)
<u>Household characteristics (<math>Z_i</math>)</u>							
Child	-0.0056 (0.27)	0.0096 (0.26)	0.0201 (0.29)	-0.0553 (0.27)	-0.0309 (0.25)	-0.1146 (0.28)	-0.0525 (0.28)
Detached, semi-detached or terraced house	Reference	Reference	Reference	Reference	Reference	Reference	Reference
Small apartment building	0.0579 (0.48)	0.0520 (0.49)	-0.0072 (0.53)	0.1032 (0.47)	-0.0209 (0.44)	-0.1189 (0.58)	-0.1040 (0.47)
Large apartment building	0.3095 (0.51)	0.3017 (0.53)	0.2638 (0.53)	0.3627 (0.48)	0.2198 (0.48)	0.2464 (0.57)	0.1668 (0.50)
High-rise building	-0.8268 (0.56)	-0.8272 (0.56)	-1.0324* (0.56)	-0.7808 (0.51)	-0.8860 (0.61)	-0.8873 (0.54)	-0.9433* (0.56)

**Table 2. Results of main regressions (ordered logit). (continued)**

Life satisfaction ( $LS_{it}$ )	Model 1	Model 2	Model 3	Model 4	Model 5	Model 6	Model 7
<u>Environmental variables (<math>A_i</math>)</u>							
Green space		-0.0018 (0.01)	0.0486*** <sup>a</sup> (0.01)				
Green space squared			-0.0007*** <sup>a</sup> (0.00)				
Distance (in 100 m)				-0.0446* (0.02)	0.1273 <sup>b</sup> (0.09)		
Distance (in 100 m) squared					-0.0104* <sup>b</sup> (0.01)		
Frequency of park visits						0.2437** (0.10)	
Park view							1.0307*** (0.26)
<u>District controls (<math>S_i</math>)</u>							
Mitte	Reference	Reference	Reference	Reference	Reference	Reference	Reference
Friedrichshain-Kreuzberg	-0.4734*** (0.12)	-0.4831*** (0.12)	-0.5027*** (0.14)	-0.4745*** (0.13)	-0.4591*** (0.14)	-0.5282*** (0.13)	-0.5141*** (0.11)
Pankow	-0.1733 (0.14)	-0.1844 (0.17)	-0.1657 (0.17)	-0.1275 (0.12)	-0.0881 (0.11)	-0.1422 (0.15)	-0.1402 (0.12)
Charlottenburg-Wilmersdorf	-0.1476 (0.09)	-0.1650 (0.14)	-0.0373 (0.16)	-0.1279 (0.10)	-0.0919 (0.10)	-0.0933 (0.11)	-0.0378 (0.10)
Tempelhof-Schönefeld	-0.0100 (0.18)	-0.0239 (0.23)	0.1729 (0.22)	0.0454 (0.18)	0.0743 (0.18)	-0.0368 (0.19)	-0.0010 (0.17)
Neukölln	-0.2473* (0.14)	-0.2477* (0.14)	-0.2149 (0.14)	-0.2466* (0.13)	-0.3010** (0.15)	-0.2255* (0.13)	-0.1600 (0.15)
Lichtenberg	0.0080 (0.17)	0.0062 (0.17)	0.0330 (0.16)	-0.0743 (0.17)	-0.0407 (0.17)	0.1181 (0.16)	-0.0113 (0.16)
Pseudo R <sup>2</sup>	0.1286	0.1286	0.1360	0.1308	0.1353	0.1337	0.1368
AIC	1125.5218	1125.4284	1116.0261	1122.6291	1116.8953	1118.9911	1115.0470
BIC	1148.0562	1147.9628	1138.5606	1145.1635	1139.4298	1141.5256	1137.5814

Cutpoints omitted. Clustered standard errors in parentheses. The number of observations is 316 for all models.

\* p<0.1, \*\* p<0.05, \*\*\* p<0.01, <sup>a</sup> jointly significant at 1% level, <sup>b</sup> jointly significant at 5% level, <sup>c</sup> jointly significant at 10% level.

Because respondents had the option to not answer single questions in the survey, the final sample consists of 316 observations. Most omissions occurred for the income variable, a potential issue we address in section 4.3.5. For all other variables, the number of omissions was relatively low so that we can safely exclude selection bias resulting from missing data points.

#### 4.2.1 Effects of socio-economic variables

The results of Model 1 reflect the standard findings of the life satisfaction literature regarding the effects of most demographic and socio-economic explanatory variables on life satisfaction (see Dolan et al. (2008) for a review). We find that income has a significantly positive effect on life satisfaction with a declining marginal effect, which is consistent with findings from other studies using cross-sectional micro data on income and life satisfaction within one country (Clark et al., 2008).<sup>19</sup> We also find the usual U-shaped relationship between age and life satisfaction (Blanchflower and Oswald, 2004). Both being married and having a partner as opposed to being single have positive effects. Moreover, self-reported health has a strongly significant, positive and monotonously increasing impact on life satisfaction.

<sup>19</sup> In addition to using net monthly individual income, we also used net monthly household income in sensitivity analyses. The results are comparable.

The impact of education is significantly positive for secondary and tertiary education compared to basic education. The effect of secondary education is slightly bigger than of tertiary education, which is in line with part of the literature (Stutzer, 2004). Being employed part-time as opposed to being employed full-time also has a significantly positive effect on life satisfaction. Previous evidence on the role of part-time and full-time work on life satisfaction is mixed (Dolan et al., 2008). A negative effect of part-time work as opposed to full-time work has been observed for men (Schoon, Hansson, & Salmela-Aro, 2005). Estimating Model 1 separately for men and woman reveals that being employed part-time has a significantly positive effect for women, while there is no significant effect for men (results not shown). We find a negative effect of being unemployed, but the coefficient is insignificant (e.g., as in Clark and Oswald, 1994), which might be because there are only 19 unemployed people in our sample.

Interestingly, we also find that respondents who have lived in a certain district of Berlin for their whole lives report significantly higher levels of life satisfaction than others. Regarding the variables that address the living environment of the respondents, we find weak evidence that the respondents' housing conditions affect their well-being. Model specifications 3 and 7 show that people living in a high-rise building are significantly less satisfied with their lives compared to those living in detached, semi-detached, or terraced houses. Regarding district effects, we find that people living in Friedrichshain-Kreuzberg and Neukölln appear to be less satisfied with their lives than people living in Mitte.

#### **4.2.2 Effects of environmental variables**

As described in section 1, there are several ways in which urban green space may influence life satisfaction. Regarding the available amount of urban green space in the living environment, Model 2 suggests that there is no significant linear effect on life satisfaction. This seems plausible because the marginal value of additional green space may depend on the current allocation. In line with this reasoning, Model 3 provides evidence of an inverted U-shaped relationship between urban green space and life satisfaction. This implies that additional urban green space first increases life satisfaction but tends to decrease life satisfaction above a certain threshold.<sup>20</sup>

The maximal positive impact of urban green space occurs at an area of 36 ha or 11.5% of the buffer area. This is well above the sample mean of 24.4 ha or 7.8%. Three quarters of the respondents in the sample have less urban green space in their living environment. Thus, measured at the means, increasing the amount of urban green space in the respondents' living environments increases their life satisfaction. The two other published studies investigating the effect of urban green space on

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<sup>20</sup> The green space coefficients are also jointly significant at the 1% level if self-reported health is excluded from the regression.

life satisfaction find a linear positive effect. However, the possibility of non-linear relationships is not considered in those papers.

In Models 4 and 5, we include the distance to the nearest urban green space greater than 5 ha as an alternative measure of the availability of urban green space in the regressions. The Euclidean distance is found to be statistically significant, with a negative effect of increasing distance (Model 4). Adding the squared distance to the model produces an inverted U-shaped relationship between distance and life satisfaction (Model 5), similarly to Model 3. The distance coefficients are jointly significant at the 5% level. Evaluated at the mean distance of 661 m, the impact is negative.<sup>21</sup>

Turning to the self-reported environmental variables, Model 6 suggests a positive relationship between the frequency of park visits and life satisfaction.<sup>22</sup> This finding can be perceived as a support of medical findings on the effect of outside activities on health. Note, however, that park visits may also depend on health and well-being, giving rise to concerns about the reverse causation and endogeneity of this variable. Note further that the frequency of park visits is related to the availability of green space and the distance to urban green spaces because both objective measures are significant predictors of park use (Schipperijn et al., 2010). Finally, Model 7 shows a strongly positive and significant effect of having a park view on life satisfaction. This suggests support for psychological findings on the positive effects of natural window views on well-being (Kaplan, 2001). However, park views could be particularly prone to a bias arising from self-selection of residential location, as a park view is an attribute that is very specific to single residences.

### **4.3 Sensitivity analysis and robustness checks**

The effects of main robustness checks are shown in Table 3. We compare the results to our preferred model specification, Model 3, which analyzes the effects of an objective green space indicator. We focus on this model because the subjective indicators used are likely to be more vulnerable to biases. In addition, Model 3 shows the best model fit as judged by the information criteria among the models using objective indicators. Robustness checks have also been carried out for the other main specifications, supporting the findings presented below.

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<sup>21</sup> The optimal distance to an urban green space greater than 5 ha is 612 m; 46.8% of the respondents live further away from an urban green space of that size.

<sup>22</sup> Results are comparable if the frequencies of summer or winter visits are used.



**Table 3. Robustness checks for life satisfaction regressions (ordered logit).**

Life satisfaction ( $LS_{ij}$ )	Model 8 Unemploy- ment	Model 9 Lifetime	Model 10 Sealing	Model 11 Sealing squared	Model 12 Status < 4	Model 13 Imputed I	Model 14 Imputed II
Log individual income	0.8608*** (0.27)		0.8716*** (0.28)	0.8987*** (0.27)	0.8289*** (0.25)	0.9006*** (0.30)	0.7611*** (0.19)
<u>Social and environmental variables (<math>A_i</math>)</u>							
Unemployment rate in planning unit	0.0361 (0.03)						
Green space	0.0452*** <sup>a</sup> (0.01)	0.0127 <sup>a</sup> (0.04)			0.0395*** <sup>a</sup> (0.01)	0.0521*** <sup>a</sup> (0.01)	
Green space squared	-0.0006*** <sup>a</sup> (0.00)	-0.0007*** <sup>a</sup> (0.00)			-0.0006*** <sup>a</sup> (0.00)	-0.0007*** <sup>a</sup> (0.00)	
Mean sealing			0.0047 (0.01)	0.0719** <sup>c</sup> (0.03)			
Mean sealing squared				-0.0006** <sup>c</sup> (0.00)			
Imputed green space							0.0349** <sup>a</sup> (0.02)
Imputed green space squared							-0.0005*** <sup>a</sup> (0.00)
Imputation dummy						0.3347 (0.75)	
Imputation dummy							-0.0742 (0.27)
Individual demographic and socio-economic characteristics	Yes	No	Yes	Yes	Yes	Yes	Yes
Household characteristics	Yes	No	Yes	Yes	Yes	Yes	Yes
District dummies	Yes	Yes	Yes	Yes	Yes	Yes	Yes
Number of observations	316	33	316	316	355	341	426
Pseudo R <sup>2</sup>	0.1366	0.1546	0.1288	0.1298	0.1393	-	-
AIC	1115.2432	111.8479	1125.2198	1123.8907	1245.2297	-	-
BIC	1137.7777	120.8269	1147.7543	1146.4251	1268.4624	-	-

Cutpoints omitted. Clustered standard errors in parentheses. \*  $p < 0.1$ , \*\*  $p < 0.05$ , \*\*\*  $p < 0.01$ ,  
<sup>a</sup> jointly significant at 1% level, <sup>b</sup> jointly significant at 5% level, <sup>c</sup> jointly significant at 10% level.

### **4.3.1 Social indicators**

We add social data on the level of the planning unit to the regression to check if the effects of urban green space shown in section 4.2 actually capture other local effects. The reason for this is that the social status of the respondents' living environments might be reflected in their endowment with green spaces. In addition to the district dummies already included, we use data on unemployment rates on the level of planning units as an indicator of social status. Model 8 shows that the effects of urban green space are comparable to those of Model 3. Therefore, even when controlling for the social status of the respondents' living environments, we find a significant effect of urban green space on life satisfaction.

### **4.3.2 Self-selection of residential location**

We further check if our environmental coefficients are biased upwards as a result of the self-selection of residential location. Individuals with higher preferences for urban green space and parks might move to greener areas of the city. As Ambrey and Fleming (2013) note, however, the evidence seems to be mixed, and several authors find that this selection bias is rather small (e.g., Chay and Greenstone, 2005). To investigate the issue further, we use the lifetime variable to split the sample and analyze the effect of urban green space only for the subsample of respondents that have been living in the same district for their whole lives. Because this subsample only includes 33 observations, we only include district dummies as control variables. The results of Model 9 again show the inverted U-shaped effect of urban green space on life satisfaction that is significant at the 1% level. This may be an indicator that the effect of urban green space expands beyond an effect of pure self-selection, even if the results have to be interpreted carefully due to changes in sample size.

### **4.3.3 Preference heterogeneity**

A further issue is that of preference heterogeneity among subsamples with differing demographic or socio-economic characteristics (Menz and Welsch, 2012; Ambrey and Fleming, 2013). One could hypothesize, for example, that families with children or people in certain age groups have stronger preferences for urban green space close to their homes, or that preferences differ with respect to gender. We follow two strategies to investigate this: i.) adding interaction effects and ii.) carrying out group comparisons (results not shown).

First, for income, age, gender, and children, we include interaction terms between these four variables with the green space variable and add them to Model 3 one by one in separate regressions. Neither of these interaction terms is significant, suggesting that there is no significant difference of the effect of urban green space on life satisfaction between the respective groups. Second, for

marital status and occupation, we perform separate analyses for subsamples.<sup>23</sup> Regarding marital status, we find that urban green space has a significant effect for those being single and for those living in a relationship. Regarding occupation, we find that the effect of urban green space is significant for the subsample of full-time employees.

Overall, there is thus no clear evidence of the systematic preference heterogeneity in our sample. This is consistent with findings of Ambrey and Fleming (2013), who also find less preference heterogeneity than anticipated. However, these results have to be treated with care, as differences in sample sizes may influence the results.

#### **4.3.4 Smaller buffer areas and alternative environmental variables**

We calculated the area of green space also available for buffers with smaller areas. Replacing the 1 km buffer with a 0.5 km or a 0.3 km buffer leads to insignificant results regarding the effect of the green space variable on life satisfaction. In these cases, the buffer areas are getting much smaller, decreasing from 3.14 km<sup>2</sup> for the 1 km buffer to 0.78 km<sup>2</sup> for the 0.5 km buffer and 0.28 km<sup>2</sup> for the 0.3 km buffer. Overall, these smaller buffer areas might be too small to capture the characteristics of a neighborhood regarding their supply with urban green space.

Regarding distance, we tried simple Euclidean, inverse and squared distances to the margin of the nearest urban green space among urban green spaces of all sizes and among urban green spaces bigger than 1 ha, 5 ha, and 10 ha, respectively. All distance measures have significant effects if the distance to the nearest green space greater than 5 ha is considered. No significant effects can be observed for the distance measures using other size thresholds.

So far, we have looked at the availability of public green space and its effect on life satisfaction. However, private green such as gardens or other green infrastructure such as street trees may also influence the life satisfaction of city inhabitants, as they can considerably alter the character of a neighborhood. Models 8 and 9 show the effects of average sealing in the respondent's living environment on life satisfaction. As observed, there is no significant linear effect. This does not change when trying out areas with different sizes over which sealing degrees are averaged. Including average sealing in a 300 m buffer and additionally including its squared value, in contrast, yields an inverted U-shaped effect that is significant at the 10% level. The effect is comparable to that of "urban green space" but significance levels are lower for sealing. These results suggest that the optimal average degree of sealing is approximately 60%. This represents a discontinuous but still quite dense urban fabric. Approximately 54% of the respondents live in areas with higher average degrees of sealing of up to 91%.

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<sup>23</sup> As in section 4.3.2, we use a smaller regression model, only including district dummies and the variable of urban green space plus its squared value to account for the smaller sample size.

### 4.3.5 Alternative samples and imputed data

To address the concern of potential selection bias due to missing data points in our sample, we ran robustness checks with larger samples and with imputed data sets. Model 12 shows how parameter estimates change if we enlarge the data set to include respondents for whom only ZIP codes are known, increasing the number of observations to 355. Urban green spaces are still highly significant for life satisfaction even though the size of the effect slightly decreases.

Given that we are faced with 50 missing observations on the income variable, we rerun the main regressions using imputed data as a robustness check. First, we only impute income data, using multiple univariate imputation with truncated regression and 50 imputations.<sup>24</sup> The results of Model 13 show that there is still a significant effect of urban green space on life satisfaction with increased effect sizes. The income coefficient also increases and is still highly significant. Second, we impute income and the area of urban green space together. To do so, we treat all observations with a location status greater than two as missing and impute also the data on green space area. We use multiple multivariate sequential imputation with chained equations and, again, 50 imputations. This increases the sample size to 426 observations. The effects of urban green space are still highly significant. Both the coefficients of income and urban green space decrease (Model 14).

### 4.4 Valuation of urban green space using the life satisfaction approach

Following the methodology described in section 3.1, we calculate the implicit MRS between life satisfaction and the environmental variables of green space area and distance, respectively. To account for the non-linear relationships, we calculate the MRS for different combinations of income and green space availability based on equations (3) or (4), depending on the specification. The implicit MRS is shown in Table 4.

Regarding green space area, the implicit MRS is EUR 25.03 per person per hectare per month based on average green space availability and average income. The MRS estimates range from EUR -30.88 for low income and high green space availability to EUR 110.63 for high income and low green space availability.

Comparing our results to the one existing study that values urban green space using the life satisfaction approach, we find that our estimates are significantly smaller. Ambrey and Fleming (2013) find an average MRS of AUD 1,168 per household per year for a 143 m<sup>2</sup> increase in green space in the respondent's collection district. This would translate to AUD 81,678 (or EUR 52,640<sup>25</sup>) per household per year for a 1 ha increase in green space. Multiplying the MRS for the average income and average green space availability we calculated by 1.9, which is the average household

<sup>24</sup> We use multiple imputations as implemented in Stata 13.

<sup>25</sup> Converted with an exchange rate of 1.5516 EUR/AUD as of December 31, 2013.

size in our sample, we find a MRS of EUR 571 per household per hectare per year, which is two orders of magnitude smaller. In the CV studies considered by Brander and Koetse (2011), the mean WTP is USD 13,210 (EUR 10,200)<sup>26</sup> per hectare per annum, while the median WTP is USD 1,124 (EUR 868), reflecting a rather skewed distribution. This is closer to our MRS estimates and considerably smaller than the estimates of Ambrey and Fleming (2013). This comparison, however, is limited because the CV studies considered by Brander and Koetse (2011) comprise all types of urban open space, including agricultural land and forests.

**Table 4. Implicit marginal rates of substitution (MRS).**

		Individual net monthly income		
		Low (587 €)	Mean (1,444 €)	High (2,301 €)
Urban green space (non-linear) (Model 3)	Low (6.1 ha)	28.22	69.43	110.63
	Mean (24.4 ha)	10.17	25.03	39.88
	High (42.7 ha)	-7.88	-19.38	-30.88
Distance (linear) (Model 4)	All	44.93	64.18	94.14
Distance (non-linear) (Model 5)	Low (228 m)	-51.98	-127.87	-203.76
	Mean (661 m)	6.63	16.31	25.99
	High (1,094 m)	65.24	160.49	255.74

Calculations based on equation (3) for Model 4 and equation (4) for Models 3 and 5. Numbers reflect the monetary equivalent (€) of a one hectare increase in the area of urban green space or a one hundred meter decrease in the distance to urban green space, respectively, per person per month. Low (high) values correspond to the mean value minus (plus) one standard deviation.

Regarding distance, Perino et al. (2013) report that the mean marginal value of proximity to a formal recreation site is GBP 150 (EUR 167<sup>27</sup>) per meter, ranging from a minimum of GBP -41 (EUR -46) to a maximum of GBP 3,348 (EUR 3,720) in the studies considered. The marginal values in these studies are all derived from one-off payments. Our estimates range from EUR -46.46 per household per meter per year to EUR 58.31 per household per meter per year and are thus considerably smaller. However, there is still no consensus on what the relationship between WTP estimated by HP and LSA is (see Ferreira and Moro (2010) for a discussion). Moreover, the studies considered by Perino et al. (2013) refer to one-off payments, while our estimates refer to the annual MRS.

<sup>26</sup> Converted with an exchange rate of 1.2949 EUR/USD as of December 31, 2011.

<sup>27</sup> Converted with an exchange rate of 0.9000 EUR/GBP as of December 31, 2009.

## 5 Discussion and conclusions

We use four individual green space measures to explore how urban green space affects the well-being of the residents of Berlin, Germany. We combine spatially explicit data derived from a customized web survey with spatially highly disaggregated GIS data on urban green spaces to carry out our analyses.

The first measure we analyze is the amount of urban green space in a certain buffer area around a respondent's residential address. This objective measure combines the notion of distance with the notion of the absolute availability of urban green space. Both aspects are significant determinants of the actual use of urban green space (Schipperijn et al., 2010). This green space measure may thus reflect the degree to which green spaces are actively used but also the degree to which they are able to mediate adverse environmental impacts. The impacts of increased outdoor activity levels may also be captured by the objective indicator "distance to the nearest green space greater than 5 ha" and in the subjective indicator "frequency of park visits," which we also analyze. Moreover, park view is included in a separate regression to see whether we can confirm the positive effects of views onto natural elements found in the psychological and medical literature.

With respect to the objective indicators, our results suggest that the effects of green space are non-linear with the marginal utility of green space first increasing and then decreasing. This is supported by the effects of distance and sealing, which also show significant non-linear, inverted U-shaped effects. One possible explanation for an inverted U-shaped relationship might be that living very close to urban green spaces may not only be associated with amenities but also with disamenities arising, e.g., from noise, congestion or fear of crime (Bixler and Floyd, 1997; Kuo et al., 1998). This issue has also been discussed in the hedonic pricing literature, suggesting that the positive effect of parks on the values of nearby properties very much depends on the quality and usage of the park (Crompton, 2001). In addition, it seems plausible that people living in urban environments not only have preferences for urban green but also have preferences for living close to infrastructure, shops, schools, or work. Studies using the LSA to investigate scenic amenities (Ambrey and Fleming, 2011) or land cover (Kopmann and Rehdanz, 2013) support the existence of non-linearities.

Further results suggest that the optimal amount of urban green space in a 1 km buffer around residential addresses is 36 ha, or 11.5% of the buffer area. Because three-quarters of the respondents have less than this amount of urban green space available in their living environments, green space is, overall, in insufficient supply in the case study area in Berlin. This also implies positive MRS estimates evaluated at the means of green space area and income. Based on mean green space availability and mean income, the implicit MRS is EUR 25 per person per hectare per month. For

city management, our results imply that policies should aim at increasing the supply of green spaces in areas where they are particularly scarce. Moreover, a more homogenous supply of urban green space should be targeted.

The analyses of the subjective green space indicators show that both the frequency of park visits and park views have positive well-being effects, with the effect of park views being particularly strong and significant. The strong effect of park views, however, may also be caused by the self-selection of the residential location. It seems plausible that park views are more prone to bias from self-selection than the variable of urban green space because the latter captures the character of a relatively large area, while a park view is an attribute that is very specific to single residences. Moreover, the effect of park visits may be biased by reverse causation, as life satisfaction itself may influence the frequency of park visits.

Despite the fact that we can show positive effects of green space on life satisfaction over a broad range of specifications and robustness checks, some issues arise that may bias the estimated income coefficients and, therefore, the estimated implicit MRS.

The endogeneity of income may be a potential issue. One reason for this is that the observed relationship between economic conditions and well-being could be explained by unobserved heterogeneity if personality traits are not considered as control variables (see, e.g., Boyce, 2010). It seems likely that on average, happier individuals tend to lose their job less often, to be re-employed more easily, or to find jobs that are better paid (see Frey and Stutzer (2002) for a discussion). This might bias the income coefficient upwards and thus bias the MRS downwards. Introducing character trait controls may reduce the potential upward bias from unobserved heterogeneity. Unfortunately, it was outside the scope of our survey to include questions about character traits so that we cannot control for unobserved individual heterogeneity in our cross-sectional data set.

Additional bias may arise from the fact that people compare their income to the incomes of other individuals (e.g., Clark et al., 2008). Because comparisons to other individuals' incomes are mostly made upwards, this effect can bias the income coefficient downwards if no appropriate controls are included in the regressions. Some studies try to instrument income to avoid potential biases (e.g., Powdthavee, 2010). However, instrumenting is not without problems either, and there is no agreement yet on how to instrument income, particularly in cross-sections.

Overall, the directions of potential biases of our income coefficient are ambiguous. Our income coefficients are larger than in studies using panel data (e.g., Levinson, 2012) or repeated cross-sections (e.g., Ferreira et al., 2012) but comparable to other cross-sectional studies (e.g., Ferreira

and Moro, 2010). Given that we report relatively high income coefficients, our MRS estimates are likely conservative lower bounds for real MRS.

Further limitations to our study arise from the fact that we cannot control for the quality of urban green spaces, which may influence well-being effects. However, currently available GIS data do not contain the necessary information, so a refinement of the analysis in this direction has to be deferred to future research.

Despite these limitations, and even though more research is needed to refine MRS estimates from studies using the LSA, our study underlines the importance of urban green space for city inhabitants. Particularly in inner-city areas, green space is often in insufficient supply. This is also the case for our Berlin case study, where three-quarters of the respondents have less than the optimal amount of green space available in their living environment. Based on our well-being analysis, increasing the average supply of urban green space and aiming at a more homogenous distribution would be preferable to current allocations.



## Appendix A – Overview of the data used for the analyses

**Table A-1. Definitions and summary statistics of demographic, socio-economic, and environmental variables.**

Variable name	Definition	Mean	Std. dev.	Min	Max
<i>Life satisfaction (LS<sub>ij</sub>)</i>					
<b>Life satisfaction</b>	Overall life satisfaction; measured on a Likert scale from 0 “very dissatisfied“ to 10 “very satisfied“	6.728	2.08	0	10
<i>Individual demographic and socio-economic characteristics (Y<sub>i</sub>, X<sub>i</sub>)</i>					
<b>Log individual income</b>	Natural logarithm of the total net monthly individual income in Euros	7.117	0.59	3.912	8.923
<b>Gender</b>	Gender dummy; 1 if “male“, 0 if “female“	0.509	0.50	0	1
<b>Age</b>	Age; measured in years	45.97	14.31	18	78
<b>Single</b>	Marital status dummy; 1 if “single“, 0 else; reference category	0.310	0.46	0	1
<b>Married</b>	Marital status dummy; 1 if “married“, 0 else	0.342	0.48	0	1
<b>Partner</b>	Marital status dummy; 1 if “living in a relationship“, 0 else	0.228	0.42	0	1
<b>Separated<sup>a</sup></b>	Marital status dummy; 1 if “separated“, 0 else	0.019	0.14	0	1
<b>Divorced<sup>a</sup></b>	Marital status dummy; 1 if “divorced“, 0 else	0.089	0.28	0	1
<b>Widowed<sup>a</sup></b>	Marital status dummy; 1 if “widowed“, 0 else	0.013	0.11	0	1
<b>Very bad health<sup>a</sup></b>	Health dummy; 1 if “Very bad health“, 0 else; reference category	0.035	0.18	0	1
<b>Bad health<sup>a</sup></b>	Health dummy; 1 if “Bad health“, 0 else	0.177	0.38	0	1
<b>Fair health</b>	Health dummy; 1 if “Fair health“, 0 else	0.275	0.45	0	1
<b>Good health</b>	Health dummy; 1 if “Good health“, 0 else	0.380	0.49	0	1
<b>Very good health</b>	Health dummy; 1 if “Very good health“, 0 else	0.133	0.34	0	1
<b>Basic education</b>	Education dummy; 1 if education on ISCED level 1 or 2 <sup>b</sup> ; reference category	0.073	0.26	0	1
<b>Secondary education</b>	Education dummy; 1 if education on ISCED level 3 or 4 <sup>b</sup>	0.453	0.50	0	1
<b>Tertiary education</b>	Education dummy; 1 if education on ISCED level 5 or 6 <sup>b</sup>	0.475	0.50	0	1
<b>Full-time employed</b>	Occupation dummy; 1 if “full-time employed“, 0 else; reference category	0.462	0.50	0	1
<b>Part-time employed</b>	Occupation dummy; 1 if “part-time employed“, 0 else	0.130	0.34	0	1
<b>Unemployed</b>	Occupation dummy; 1 if “unemployed“, 0 else	0.060	0.24	0	1
<b>Unable to work</b>	Occupation dummy; 1 if “unable to work“, 0 else	0.038	0.19	0	1
<b>Retired</b>	Occupation dummy; 1 if “retired“, 0 else	0.155	0.36	0	1
<b>Student</b>	Occupation dummy; 1 if “student“, 0 else	0.092	0.29	0	1
<b>Other occupation</b>	Occupation dummy; 1 if “other occupation“, 0 else	0.060	0.24	0	1
<b>Migration background</b>	Dummy variable; 1 if at least one parent of the respondent is of a nationality other than German, 0 else	0.335	0.47	0	1
<b>Lifetime</b>	Dummy variable; 1 if respondent has been living in the same district for her whole life	0.104	0.31	0	1

**Table A-1. Definitions and summary statistics of demographic, socio-economic, and environmental variables. (continued)**

<i>Household characteristics (Z<sub>i</sub>)</i>					
<b>Child</b>	Dummy variable; 1 if at least one child under the age of 12 is living in the household, 0 else	0.111	0.31	0	1
<b>Detached, semi-detached, or terraced house</b>	Housing dummy; 1 if “detached, semi-detached or terraced house“, 0 else; reference category	0.070	0.25	0	1
<b>Small apartment building</b>	Housing dummy; 1 if “apartment building with 3 to 8 apartments“, 0 else	0.294	0.46	0	1
<b>Large apartment building</b>	Housing dummy; 1 if “apartment building with 9 or more apartments (but no high rise)“, 0 else	0.541	0.50	0	1
<b>High-rise building</b>	Housing dummy; 1 if “high rise“, 0 else	0.095	0.29	0	1
<i>Environmental variables (A<sub>i</sub>)</i>					
<b>Green space</b>	Hectares of urban green space in a 1 km buffer area around the respondent’s home	24.43	18.26	1.094	98.120
<b>Distance</b>	Distance to nearest urban green space > 5 ha; measured in metres	660.84	432.6	0.5	2495.3
<b>Frequency of park visits</b>	Frequency of park visits; measured on a 6-point scale from 0 “never“ to 5 “(almost) daily“	1.932	1.20	0	5
<b>Park view</b>	Dummy variable; 1 if respondent has a view of a park from his/her home, 0 else	0.130	0.34	0	1
<i>District controls (S<sub>j</sub>)</i>					
<b>Mitte</b>	District dummy; 1 if respondent lives in this district, 0 else; reference category	0.165	0.37	0	1
<b>Friedrichshain-Kreuzberg</b>	District dummy; 1 if respondent lives in this district, 0 else	0.092	0.29	0	1
<b>Pankow</b>	District dummy; 1 if respondent lives in this district, 0 else	0.155	0.36	0	1
<b>Charlottenburg-Wilmersdorf</b>	District dummy; 1 if respondent lives in this district, 0 else	0.136	0.34	0	1
<b>Tempelhof-Schöneberg</b>	District dummy; 1 if respondent lives in this district, 0 else	0.187	0.39	0	1
<b>Neukölln</b>	District dummy; 1 if respondent lives in this district, 0 else	0.114	0.32	0	1
<b>Lichtenberg</b>	District dummy; 1 if respondent lives in this district, 0 else	0.152	0.36	0	1

The number of observations included is 316 for all variables.

<sup>a</sup> In the regression analyses in section 4, the dummy variables “separated“, “divorced“, and “widowed“ as well as “very bad health“ and “bad health“ are merged into one variable respectively due to the low number of observations.

<sup>b</sup> For an overview of how the German educational achievements translate into the internationally comparable ISCED levels see Statistisches Bundesamt (2010).

## **Appendix B – Survey details**

The survey was implemented using a pre-selected web panel with approximately 100,000 registered members throughout Germany and 4,000 members in Berlin aged 18 years or above. It was implemented and executed via a professional polling agency. Potential survey participants were invited via email to participate in the survey. In this email, a link to the survey was provided, but the topic of the survey was not further specified. Having clicked on the link, potential participants were directed to a standard starting screen, which clarified the expected length of the survey and the potential reward to be gained. The topic announced on this starting screen was kept very general, only indicating that the survey was related to city life, to avoid self-selection in terms of interest in environmental issues. In total, the questionnaire consisted of between 25 and 45 questions, depending on whether the respondents regularly visited parks or not.

An early version of the questionnaire was pretested with university students. The final version of the survey was pretested with a set of 50 participants. Responses to the survey were checked according to several quality criteria, including the time taken to complete the survey and obvious answer patterns that revealed that respondents had clicked through the survey without paying attention to questions and answers. Approximately 10% of the observations have thus been eliminated from the sample. In addition, a thorough validity test was carried out, checking answers for obvious inconsistencies in content. This led to a further exclusion of 25 observations from the sample.

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